

Disturbance-generated niche-segregation in a structured metapopulation model

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ABSTRACT

In studying coexistence that is maintained by local disturbances via a trade-off between fecundity and competitiveness, we use the theory of structured metapopulations to consider the local- and metapopulation-scale dynamics on the same footing. In contrast to the predictions of the usual patch-occupancy models, our study reveals that this kind of diversity, while certainly possible, is severely constrained by the requirement of partitioning of the continuum of patches with different population densities, i.e., patches of different ages. Adaptive dynamics, evolutionary stable coalitions and evolutionary branching are investigated in the model. We conclude that the partitioning of the patch-age, as a niche dimension, does not differ in any essential respect from partitioning of a resource-axis.

Subject headings: Diversity, disturbance, niche, metapopulation

1. Introduction

Maintenance of species diversity via disturbance (Connell 1978; Hastings 1980; Huston
2 1979, 1994) is a central issue of ecology. In the most commonly considered case, it is
assumed that the ability to colonize and/or exploit an empty habitat can be increased at the
4 cost of decreasing local competitiveness. In a constant environment, the better competitors
(the “ K -strategists”) outcompete the good colonizers/exploiters (the “ r -strategists”) in
6 each habitat, so the latter ones disappear. If, however, the local sub-populations are
destroyed regularly, the r -strategists may always have a chance to exploit a newly emptied
8 habitat before they are expelled from the previous one. This way, the long-term coexistence
of the r - and K -strategist (or “early-successional” and “late successional” species, Rees
10 et al. 2001) is maintained.

Such mechanism requires a metapopulation (Levins 1968) structure, i.e., a collection of
12 local populations connected via dispersal. The usual way of modelling is based on a serious
simplification: a patch is either empty, or fully occupied by a single species (e.g. Nee and
14 May 1992; Amarasekare et al. 2004). Occupied patches become empty due to catastrophes.
Empty patches become occupied, when colonized by dispersers from the occupied patches.
16 In this context, better competitive ability is modelled via the possibility of overcolonization
of the less competitive species. These “patch occupancy” models can also be interpreted
18 outside the patch disturbance framework, as one can define the “patch” as a single “safe
site”, possibly occupied by a single individual (Tilman et al. 1994; Calcagno et al. 2006;
20 Geritz 1995; Geritz et al. 1999; Kisdi and Geritz 2003).

As we are interested in disturbance-mediated coexistence, we stick to the original
22 “population of populations” interpretation of metapopulation modelling. In this context,
overcolonization is a process in which the more competitive species gradually ousts the
24 other one from the patch. This succession may take many generations. As the traditional

metapopulation models do not follow the local population dynamics, the details of the
26 real process of colonization and competition remain hidden. Here we look for a deeper
understanding via following the population growth and the process of competitive exclusion
28 within the patches explicitly, i.e., by considering the local- and metapopulation-scale
dynamics on the same footing. To this end, we study the phenomenon in a *structured*
30 metapopulation model (Gyllenberg and Metz 2001; Metz and Gyllenberg 2001; Parvinen
and Egas 2004; Parvinen 2006).

32 There are two basic ways of being an r -strategist (Pacala and Rees 1998). Either, it
can be a good colonizer that arrives first to the empty patch. This is the widely investigated
34 case of competition-colonization trade-off (Hastings 1980; Nee and May 1992; Tilman
et al. 1994; Calcagno et al. 2006; Kinzig et al. 1999). Or, the r -strategist can exploit the
36 resource-rich environment of the new habitat faster, than the K -strategists (successional
niche segregation, Pacala and Rees 1998; Szabó and Meszéna 2007). The latter type of
38 trade-off is our choice for the current investigation.

2. Model definition

40 We adapt our model from the structured metapopulation model of Parvinen (2006)
by introducing the trade-off, that decreased fecundity is the cost of increased within-patch,
42 or local, competitiveness. The dispersal properties are the same for all variants. An infinite
number of habitat patches is assumed. For numerical simplicity, discrete, non-overlapping
44 generations are used. Individuals reproduce once and die afterwards; simulation steps
follow the generations. In each step, after reproduction, a fraction d of the newborns enters
46 the dispersal pool; the rest remains in their original patch. Dispersers survive to the next
step with probability k and immigrate into a randomly chosen patch with probability α .
48 In any patch, the immigrants and the non-dispersed individuals together form the parent

population for the next reproduction. Catastrophes may occur randomly in any patch
 50 after the dispersal process. The probability μ of a catastrophe is independent of the local
 population size. A catastrophe kills all individuals in the patch. Nevertheless, the patch
 52 remains habitable, and can be re-colonized by dispersers from the disperser pool.

The fecundity $f(s, \sum N)$ of an individual depends on its inherited “strategy” s that
 54 characterizes the species, and on the total density $\sum N$ of the individuals in the patch. The
 latter dependence is of the Ricker-type (Ricker 1954)

$$f\left(s, \sum N\right) = f_0(s) \exp\left(-\frac{\sum N}{K(s)} \ln f_0(s)\right),$$

56 where $f_0(s) = f(s, 0)$ is the fecundity at zero density. The carrying capacity $K(s)$
 is the equilibrium density, when the strategy is alone in an isolated patch. (That is,
 58 $f(s, K(s)) = 1$) for all s .) It can be considered as the measure of the local competitiveness: In
 absence of catastrophes, species with the largest K outcompetes all others from all patches.

60 The strategy $s \in [0, 1)$ characterizes the species along the r - K continuum. We assume
 the dependences $K(s) = (1 - s)^{1/\beta}$ and $f_0(s) = 1 + \gamma s^{1/\beta}$. The value $s = 0$ corresponds to
 62 the extreme K strategy that maximizes the carrying capacity; $K(0)$ is scaled to 1. When s
 approaches 1, fecundity increases, while the carrying capacity goes to zero. This end of the
 64 scale represents the r strategy. Parameter β scales the shape of the trade-off (Fig. 1). It is
 concave for $0 < \beta < 1$ and convex for $\beta > 1$. Parameter γ determines the maximal possible
 66 value of fecundity.

3. Results

68 Fig. 2 illustrates coexistence of four different species. In the left panel their abundances
 $N_s(t)$ in a single patch are plotted as a function of the patch age t since the last catastrophe.

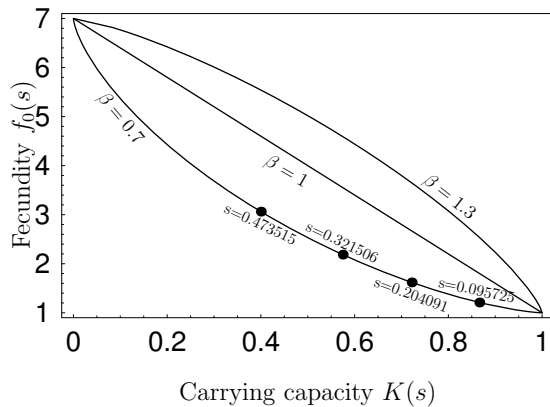


Fig. 1.— Trade-off between fecundity $f_0(s)$ and local carrying capacity $K(s)$ for the parameter values $\beta = 0.7$ (concave trade-off), $\beta = 1$ (linear), and $\beta = 1.3$ (convex); $\gamma = 6$. The four dots on the concave curve represent the four strategies in Fig. 2.

70 With the given parameters, the local population dynamical processes are much longer
 than the generation time, so we have reasonably smooth time-courses. The probability
 72 that a patch survives until a given age is plotted as a thick curve. On the right panel
 one can observe that, initially, the species with the largest colonization ability has the
 74 highest fecundity. It loses its advantage and is taken over by another species with lower
 colonization, but higher competitive ability, when the total size of the local population
 76 increases and the local competition increases. This pattern of succession is repeated until
 the species with the largest competitive ability reaches its carrying capacity.

78 How many strategies can actually coexist in the metapopulation? In particular, would
 it be possible to invade the established coalition by a new species? To tackle these issues we
 80 need a notion of fitness. In general, fitness is the long-term exponential growth rate of an
 invader in an environment set by the resident coalition (Metz et al. 1992). A newly arrived
 82 strategy invades if, and only if, its fitness is positive. The fitness is zero for an equilibrium
 population. In the metapopulation setting we find it more convenient to study fitness as
 84 the logarithm of the basic reproduction ratio R_m of the *dispersal* generations. R_m is defined

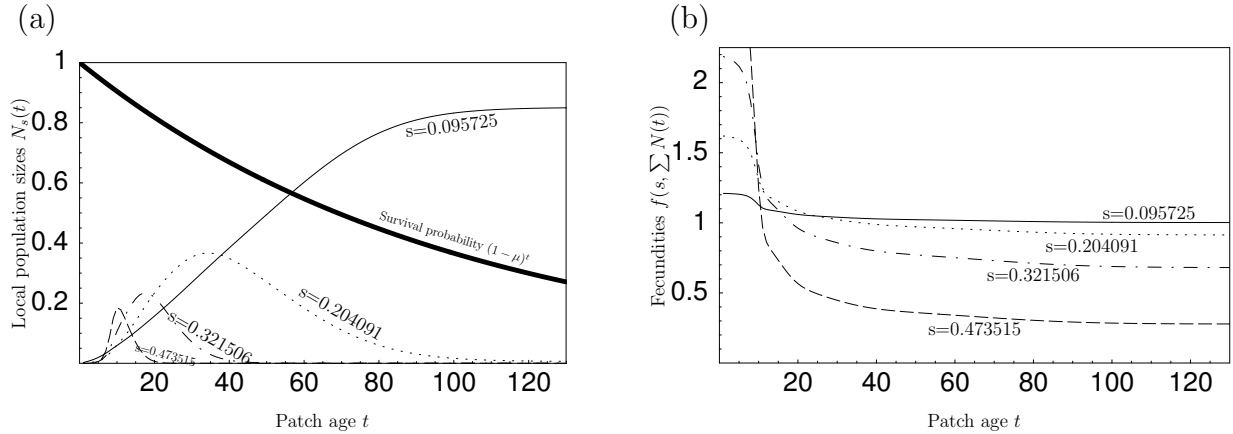


Fig. 2.— Coexistence of four different strategies. Local population size (a) and fecundity (b) of the strategies as a function of patch age. The thick curve on panel (a) represents the survival probability of the patch until the specified age. Note that the strategy with the highest K dominates the old patches. However, only a fraction of the patches survives so long. (Parameters: $\beta = 0.7$, $\gamma = 6$, $d = 0.005$, $k = 0.95$, $\alpha = 0.8$, $\mu = 0.01$. The same values are used in the other figures when not specified otherwise.)

as the expected number of dispersers produced by the clan initiated by a mutant disperser
 86 in the environment set by the resident (Gyllenberg and Metz 2001; Metz and Gyllenberg
 2001; Parvinen 2006). (Do not confuse dispersal generation with the individual generation.
 88 The former one constitutes a variable number of the latter one.) The (sign) equivalence
 of the two fitness measures is a non-trivial result (Gyllenberg and Metz 2001; Metz and
 90 Gyllenberg 2001).

Fig. 3a presents the fitness function when the coalition of Fig. 2 is established. The
 92 four established strategies have zero fitness, as expected. The non-trivial feature is that all
 other strategies have negative fitness. It means that the four strategies chosen for Fig. 2
 94 form an uninvadable, i.e., an evolutionarily stable coalition.

How does one reach such an uninvadable strategy coalition? One option is to assume

96 that all strategies (with a discretization) are present initially. Then one can follow the
metapopulation dynamics and see which strategies remain, as illustrated in Fig. 3b. (Here
98 we iterated the dispersal generations using the calculated R_m values for each strategy. As
the dispersal generations are of unequal length, this simulation does not exactly match the
100 real time-course. Nevertheless, the final state is unaffected.) We observe that the strategies
very soon form several clusters. The width of these clusters becomes narrower in time.
102 Except of the effect of the discretization, the end result is the uninvadable coalition of Fig.
2.

104 Fig. 3c represents the possibility of evolutionary emergence of an uninvadable coalition
starting from a single strategy. Here we iterated the real local population dynamics
106 (reproduction and dispersal) and random catastrophes in 1000 patches. Every 1000th
generation was considered as a single evolutionary step. In each of these steps a single
108 random mutation occurs. (A strategy slightly different from a randomly chosen present
strategy is introduced with a small population size. Mutation step size follows a Gaussian
110 distribution with standard deviation 0.01.) Repeated mutations and invasions result in the
change of the strategies constituting the coalition. One can observe several consecutive
112 events of evolutionary branching (Geritz et al. 1997, 1998). The evolutionary process ends
up in the same uninvadable coalition again, now with a cloud of mutants. Note, however,
114 that the uninvadable coalitions obtained with the two described methods are not necessarily
always equal.

116 Fig. 4 depicts the parameter-dependence of the final coalition obtained starting from all
strategies (with a discretization). We observe that coexistence of four strategies is common;
118 coexistence of five ones is restricted to narrow parameter ranges. The β dependence shows
that, given the other parameters, the investigated trade-off curve with $\beta = 0.7$ is around
120 to be optimal to maintain diversity. If the disturbance rate is high, i.e., the average

patch lifetime $\frac{1}{\mu}$ is short, only the extreme r strategist survives. Longer patch ages allow
122 coexistence such, that the average strategy of all individuals moves to the direction of K
strategy. At low disturbance rate the most competitive species becomes dominant, as one
124 can observe from the change of the average strategy.

Evolutionarily stable coalition of more than five strategies was not observed. As we
126 are not aware of, or hope for, an analytic way of proving an upper bound of the number
of coexisting strategies, we cannot exclude the possibility of model versions supporting a
128 slightly larger coalition. Nevertheless, the limited possibility maintaining high diversity this
way is a clear conclusion.

130

4. Discussion

Understanding the factors shaping species diversity is a fundamental problem with
132 great practical relevance for nature conservation. The role of environmental fluctuations
is an important focus of investigation. Here we analysed the possibility of coexistence
134 maintained by a trade-off between fecundity and local competitiveness via a structured
metapopulation model. The novelty of our approach is that we follow the local as well as the
136 metapopulation-scale dynamics. This way we acquired a clearer picture on the possibility
of disturbance-related coexistence than the one was provided by the patch-occupancy
138 models. We found that coexistence is based on, and constrained by, the possibility of the
successional segregation along the patch-age axis. Evolutionary stability of the coexisting
140 coalition was analysed.

Applicability of limiting similarity (MacArthur and Levins 1967; Abrams 1983) was
142 noted already in the patch occupancy models. However, one is free to choose the parameters
describing overcolonization in that framework. Therefore, coexistence of an arbitrary

144 number of species can be arranged (Tilman et al. 1994). Augmentation of the picture with
the possibility of back-colonization results in further constraining the coexistence of similars
146 (Kinzig et al. 1999). Still, with proper choice of the parameters, one can arrange arbitrarily
large coalitions. This behavior is in line with the fact that the patch occupancy models
148 are isomorphic to the asymmetric version of the Lotka-Volterra competition model (Kisdi
1999). Existence, or non-existence of back-colonization corresponds to the non-extreme, or
150 extreme asymmetry of competition, respectively (see Geritz 1995; Geritz et al. 1999 for this
distinction within non-Lotka-Volterra site occupancy models).

152 When mechanistic details of the local replacement process are factored in, as in the
current investigation, the possibilities are more constrained. The expected patch life-time
154 should be long enough to allow several successional steps that cannot be speeded up
arbitrarily. However, longevity of the patches is detrimental for the r -strategists: The lower
156 rate of disturbances decreases the ratio of the fresh patches needed for them. Therefore, in
agreement with the intermediate disturbance hypothesis (Connell 1978), neither the low,
158 nor the high level of disturbance supports significant diversity. In between, for intermediate
levels of the catastrophe rate, coexistence of several, but not very many, successional states
160 were observed.

It is also possible to model the successional replacement process within a single patch
162 (e.g. Fig. 6.1 in Tilman 1988, p. 187). However, if a single habitat of finite lifetime is
considered separately, the long-term survival of the species at the metapopulation scale
164 is defined outside the scope of the study. In this case one has to *postulate* the set of
species, that has the possibility to enter the habitat. To see the complete picture, parallel
166 consideration of the local and the global scales was essential. On the one hand, the global-
scale competitive exclusion determines the species composition of the metapopulation.
168 On the other hand, study of the local operation of successional replacement provides the

constraints for the global-level processes. See Szabó and MeszÉna (2007) for another study
170 of this kind.

The presented results are consistent with the general impression about the real-word
172 processes. In particular, our Fig. 2 is very reminiscent to the empirically observed succession
(Fig. 8.22 of Tilman 1988, p. 284).

174 Our investigation is inherently related to a major debate of community ecology. Niche
theory (Hutchinson 1978; Leibold 1995; Chase and Leibold 2003), a centrepiece of classical
176 ecology, asserts that competitive exclusion operates between species attempting to occupy
the same niche, so the availability of the different niches constraints diversity. In contrast,
178 the theory of disturbances (Connell 1978; Hastings 1980; Huston 1979, 1994) argues that
repeated disturbances in a non-equilibrium ecosystem are able to increase diversity beyond
180 the prediction of the equilibrium theory via weakening competition.

The controversy about, and the axiomatic status of, the competitive exclusion principle
182 (“complete competitors cannot coexist”) was reviewed already in the seminal paper of
Hardin (1960). While the principle played a key role in Hutchinson’s concept of niche
184 (Hutchinson 1978), he also introduced the doubt that the non-equilibrium situations may
be entirely different (“paradox of the plankton”, Hutchinson 1961). These distinctions
186 between equilibrium and non-equilibrium, niche and disturbance are used to be regarded as
fundamental (Huston 1979, 1994).

188 However, Hardin’s argument about the unavoidability of niche segregation was
reformulated for fluctuating situations by Chesson (1991). A general theory of fluctuation-
190 mediated coexistence was developed on this basis (Chesson 1994, 2000*a,b*). The fallacy
of the idea, that “disturbance” decreases “competition” somehow in a general way,
192 was discussed in detail by Chesson and Huntly (1997). In line with the emergence of
this new synthesis, the disturbance-maintained coexistence is now also considered as a

194 niche-segregation along life-history trade-offs (Rees et al. 2001; Amarasekare 2003).

Our model is a clear demonstration of the non-existence of a principal difference
196 between equilibrium and non-equilibrium ecology. We considered the very ecological
situation which is the archetype of the fluctuation-maintained diversity. Nevertheless, the
198 same pattern can also be seen as an equilibrium one. At the metapopulation level, the
dynamical variables of the model are the $N_s(t)$ distribution functions. For each strategy,
200 it specifies the distribution of the population between the patches of different ages. In
terms of these variables, the dynamics is deterministic and converges to the equilibrium
202 distribution displayed in Fig. 2.

From the local point of view, the less competitive r -strategist species are able to survive
204 temporarily because catastrophes transitionally reduce local competition. However, the
global situation cannot be described as if “competition”, in general, would be weaker than
206 in equilibrium. It is neither justified to regard the r -strategist as a weak competitor on the
metapopulation scale. As both kinds of simulations (Figs. 2b and 2cd) demonstrate, the
208 metapopulation-level competition is reduced only between sufficiently different strategies
that present in different periods during the succession process. Competition remains
210 strong against the intermediate strategies. This is exactly the picture of niche-segregation:
Coexistence is made possible by the reduced between-species competition. In turn,
212 weakening of competition is caused by adaptations to different ecological roles. In our case,
these roles are the presence of the species at the different successional stages. That is, the
214 essential issue is the partitioning of the patch age, as a niche dimension. An r -strategist
species, if survives, is the best competitor in its own niche.

216 A general theory of niche-segregation, formulated for equilibrium, was proposed
elsewhere (Meszéna et al. 2006, see also Gyllenberg and Meszéna 2005). It is based on
218 the notion of the “regulating variables”, i.e., on the environmental variables involved in

the feedback loop of population regulation, like resource concentrations. The theory states
220 that coexisting species should differ in their relation to these variables. More specifically:
they should differ both on their (differential) impact on, and sensitivity towards, the
222 regulating variables. Similarity in these respects restricts coexistence to a narrow range of
the parameters, i.e., makes it unlikely.

224 Accordingly, the set of regulating variables constitutes the niche space to be partitioned
between the species (Meszena et al. 2006). In case of a resource continuum (e.g., a
226 continuum of food size), this set is a continuous one: the resource concentrations for all
food-sizes have to be considered as separate regulating variables. Then the theory reduces
228 to the well-known picture of resource partitioning. While there is no absolute lower bound
of similarity for the coexisting populations, as a rule of thumb, the usual idea of niche
230 segregation prevails (Szabo and Meszena 2006; Barabas and Meszena, *in prep.*).

This equilibrium theory applies to our model without any caveat. Here the total
232 density $\sum N(t)$ for the patch-age t plays the role of regulating variable. It describes the
resource exploitation, i.e. the strength of competition in the patches of age t . If this
234 quantity is known for all patch-ages, then the fecundities of the different strategies in any
patch are calculable. This way, we have a continuum of the regulating variables along the
236 patch-age axis. It is in a complete analogy with the continuum of resource densities (more
exactly: the exploitations thereof) in the food-size case.

238 In this study we opted to keep the dispersion parameters constant. In reality, one can
expect the dispersion rate to be optimized at a higher value for an r -strategist, than for
240 a K -strategist. Moreover, the higher migration rate itself (even without higher fecundity,
as in the pure case of competition-colonization trade-off, cf. Pacala and Rees 1998) can
242 contribute to the r -strategist nature of a species. In general, one should consider a trade-off
between three variables: fecundity, dispersal and competitiveness. However, the mathematical

244 structure would remain similar to our one. If the populations within a single patch are
regulated by a single factor, then the whole metapopulation is regulated by the patch-age
246 continuum of this factor. Coexistence in such a system is an issue of niche-segregation along
the patch-age axis.

248 The issue of the optimal value of dispersion parameters is related to the evolutionary
origin of the r strategy. Tilman (1988) considers fast vegetative growth, which may, or
250 may not be coupled to fast reproduction, as an adaptation to high rate of biomass loss e.g.
by herbivory. Such r strategists need not have a high dispersion rate, and will not have it
252 because of the trade-off between migration ability and the other life-history parameters.
Still, they will dominate the early stages of succession if they happen to find their way
254 to the place. In contrast, if the early stages of the succession in a metapopulation with
repeated patch catastrophes is the typical niche of an r strategist, then it is necessarily
256 selected for higher migration rate.

Note, that the term “disturbance” is not always defined with clear distinction between
258 a constantly high loss rate and repeated patch catastrophes. While both situations select
for high growth rate, only the second one is able to maintain diversity by providing a new
260 opportunity for niche segregation. A constant high rate of biomass loss does not introduce
any new regulating variable. This is, again, in line with the conclusion of Chesson and
262 Huntly (1997) that “harshness” of the environment, alone, neither decreases competition,
nor increases diversity.

264 In our model, the abiotic environment remains constant between the catastrophes.
It is like a secondary succession, when fertile soil is available from the very beginning.
266 In case of primary succession, one should take into account the facilitating effects of the
populations. Then, an additional variable (e.g. soil nitrogen level, cf. Tilman 1988, p. 214)
268 should be followed as a function of the patch age. Still, age remains a full characterization

of a patch. Therefore, the picture of niche-segregation along the patch-age axis would
270 survive. If the “resource ratio hypothesis” of the primary succession (Tilman 1988, p. 217)
is to be modelled explicitly, then one should consider two local regulating variables (e.g.
272 soil nitrogen level and light intensity on the soil surface) instead of a single one, like the
total density used here. Then, at the metapopulation level, the set of regulating variables
274 consists of the two patch-age-continuums of the two local regulating variables. All of such
details are completely overlooked by the patch occupancy models.

276 The *complete* extinction at patch catastrophe was an essential ingredient of our model.
This assumption ensures that patch-age (i.e., the time spent since the last catastrophe in the
278 patch) is a full descriptor of the state of the patch. The possibility of partial catastrophes
would require considering the whole history of disturbances at any localities. Niche analysis
280 of such situation is beyond our scope here.

It is an ongoing controversy whether evolutionary branching, that occurs in non-genetic
282 models, can be interpreted as a minimal representation of (the ecological aspects of)
speciation (see Dieckmann et al. 2004; Gavrillets 2005 for the opposing views; Dieckmann
284 and Doebeli 1999; Pennings et al. 2008 for model studies). Note also that species coexisting
along the r - K continuum are often not closely related to each other. Therefore, their
286 evolution cannot be described solely in our context. Still, the observed branching pattern
demonstrates in principle, that the relation between ecology and evolution is the same for
288 the disturbance-maintained and for the resource-competition-maintained diversity. In both
cases, the niche-theoretic “decreasing competition with increasing difference” translates to
290 the possibility of disruptive selection and evolutionary branching (cf. Christiansen 1988
for the case of resource competition). The Darwinian picture (Darwin 1859, p. 76), that
292 competition between the similar kinds drives evolution, while the decreased competition
between the differing species allows them to coexist, prevails. In contrast, the status of

294 natural selection would be unclear under a kind of general weakening of competition by
disturbance. This way, evolutionary considerations reinforce the deep connection between
296 disturbance and niche.

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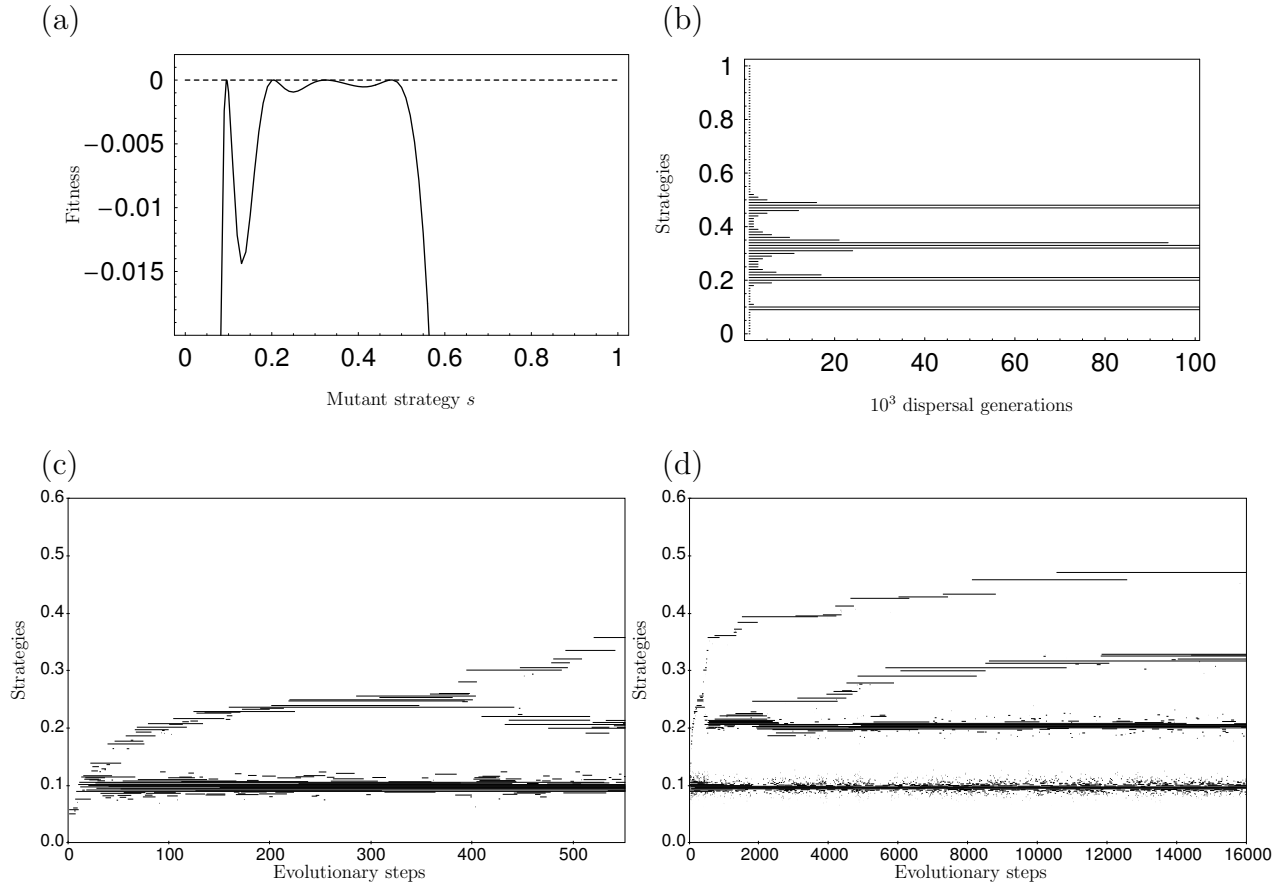


Fig. 3.— Evolutionary considerations. (a) Fitness as a function of the strategy provided, that the coalition of Fig. 2 is present in the metapopulation-level equilibrium. Observe that the four strategies forming the coalition are local fitness maxima; the maximum is equal to zero by the equilibrium condition. All other strategies have negative fitness, so the coalition is an evolutionary stable one. (b) Simulation initiated with 101 strategies $0, 0.01, 0.02, \dots, 0.99, 1$. Most of the strategies die out soon. The surviving coalition is essentially the same, as was presented in Fig. 2. Evolutionary simulation with mutation (c and d) starting from a single strategy plotted on two different time scales. Again, essentially the same coalition emerges.

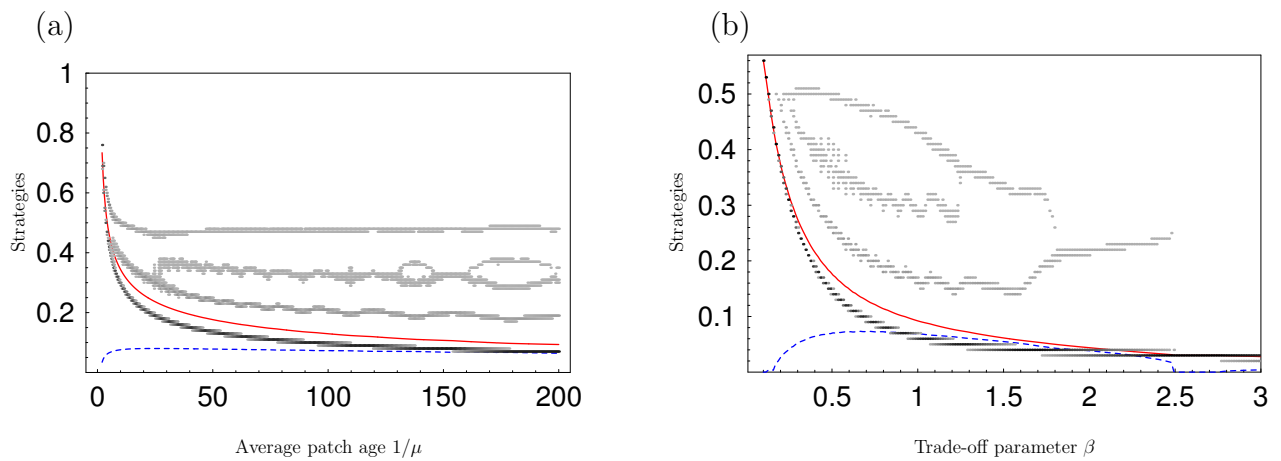


Fig. 4.— Parameter dependence of the surviving coalition in a simulation initiated like in Fig. 3b. Darkness represents abundance in the coalition. (a) Dependence on average patch age $1/\mu$. (b) Dependence on the trade-off parameter β . The average overall strategy is plotted with a thin solid curve, and the standard deviation of the strategy distribution with a dotted curve.

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